

# Indoor Particulate Concentration during Biomass burning in Central India

Rameshwari Verma<sup>1,2,\*</sup>, Khageshwar Singh Patel<sup>2</sup>, Santosh Kumar Verma<sup>1,2,\*</sup>  
Eduardo Yubero Funes<sup>3</sup>, Xiujian Zhao<sup>1</sup>

1. State Key Laboratory of Silicate Materials for Architectures, Wuhan University of Technology, No. 122, Luoshi Road, Wuhan 430070, P. R. China
2. School of Studies in Chemistry, Pt. Ravishankar Shukla University, Raipur-492010 (C.G.), India
3. Universidad Miguel Hernandez-Elche, Division de Física Aplicada, Dpto. Física y Arquitectura de Computadores, Av. Del Ferrocarril s/n. Edificio Alcudia, Spain

**Date of submission:** 20 Sep 2016, **Date of acceptance:** 22 Feb 2017

## ABSTRACT

Indoor air particulate (PM) exposure is several folds more dangerous than outdoor air owing to burning of different materials. Burning biomass emits toxic fumes that are found to be associated with numerous health problems such as respiratory diseases, etc. In our study area, approximately 80% of the population of Chhattisgarh state, central India use biomass such as wood, and cow dung as a primary source of domestic energy and therefore require proper study about indoor emission. Thus, the PM<sub>10</sub> and associated eight ions i.e. Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup>, Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup> and Ca<sup>2+</sup> from the burning of wood and cow dung in indoor in Raipur, Chhattisgarh, central India is investigated. The highest mean concentration of PM<sub>10</sub> (17697 µg m<sup>-3</sup>) and the sum of eight ions, Σion8 (38.4 mg m<sup>-3</sup>), were found from the burning of wood. The indoor concentration of PM<sub>10</sub> exceeds the guidelines levels. The wood like *Mangifera indica* emits the highest concentration of PM<sub>10</sub>. However, *Acacia arabica* is found to be acidic in nature. Thus, this result helps us to be aware of the adverse effects of indoor emission from burning. Therefore, the improved models, alternative for energy source and sufficient ventilation are supposed to be recommended option for the future.

**Keywords:** Indoor air, particulate, ions, wood, cow dung.

## Introduction

Biomass is a primary energy source in India.<sup>1</sup> India is an agricultural country with 770 million people (70% population) living in villages or rural area, and depend on various energy sources such as biomass i.e. wood, cow dung etc. for cooking, water heating and lighting in house. About 90% of rural houses and 32% of urban houses use biomass for cooking and only 25% use cleaner gases. In rural India, 62% houses use wood and 14% use cow dung while, 13% use crop residues. Many rural households have energy-inefficient clay-stoves that

produce high exposure levels of respirable particulate (PM).<sup>2</sup> Approximately 76% of all global particulate air pollution occurs indoors in the developing world.<sup>3</sup> Indoor concentration of the particulate usually exceed guidelines levels (24 hours mean PM<sub>10</sub> level of 150 µg m<sup>-3</sup>) in the range of 300 – 3000 µg m<sup>-3</sup> and may reach 30,000 µg m<sup>-3</sup> or more during the periods of cooking.<sup>4</sup> The low income families, who cannot afford modern fuels (like electricity, coal, kerosene, LPG and solar energy), rely on biomass as a primary source of fuel for cooking and water heating purpose. The low income is one of the main barriers of increasing the harmful PM concentration in indoors during biomass burning. The wood and cow dung are the highest contributor fuel in rural areas, while cleaner fuel LPG is the one in urban areas. Therefore, rural households produce greater indoor air pollution than the urban households in India as a result of the use of wood and cow

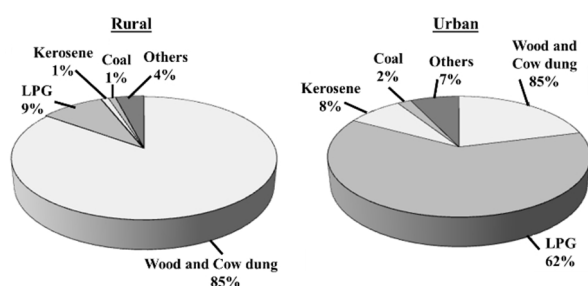
✉ Rameshwari Verma  
rbaghel9@gmail.com

Santosh Kumar Verma  
vermasantosh08@gmail.com

**Citation:** Rameshwari V, Khageshwar SP, Santosh KV, Eduardo YF, Xiujian Z. Indoor Particulate Concentration during Biomass burning in Central India. J Adv Environ Health Res 2017; 5(1):1-9

dung as major energy sources (Figure 1).<sup>5</sup> This is a challenge for providing convenient fuels until certainty that the energy services is reasonable for the rural poor families. There is a task for policy interventions and government support.<sup>6</sup>

Particulate emitted from wood burning contain OC, BC, nitrate, ammonium, sulphate, organic components and water-soluble potassium etc.<sup>7</sup>  $K^+$  has been used in several cases such as a trace element for the qualitative identification of biomass burning.<sup>8</sup> Nitrogen<sup>9</sup>, sulphur<sup>10</sup>, and halogen-containing compounds<sup>11</sup> in the PM are the function of fuel composition and combustion conditions. The principal water soluble ions emitted from biomass burning are  $Cl^-$ ,  $NO_3^-$ ,  $SO_4^{2-}$ ,  $NH_4^+$  and  $K^+$ . Cow dung emits numerous amounts of  $NO$ ,  $NH_3$ ,  $SO_2$  and  $CH_3Cl$  because of their high N, S and Cl content.<sup>7</sup> It was reported that water soluble ions such as sulfate, nitrate and other acid rain related pollutants have severe effects on human health.<sup>12</sup> The continuous use of wood for cooking is an unavoidable issue in developing countries and also in India, owing to its adverse health effect.<sup>13,14</sup>



**Fig. 1** Distribution of cooking fuels for rural and urban households in India (NSSO, 2010, Ref. 4).

Women and children are exposed especially to respiratory infections i.e. acute and chronic respiratory diseases (particularly obstructive pulmonary disease) caused by indoor air pollution, because they spend most of their time at home during fuel burning. The health problems include acute lower respiratory infections in children up to 5 years, chronic obstructive pulmonary disease in adults, tuberculosis, asthma, cataracts, among others.<sup>15,16</sup> In developing countries, indoor air

pollution is more than 5-times greater than the outdoor air pollution causing adverse health effects.<sup>17,18</sup> The kitchen used for cooking in various Indian households is found to be poorly ventilated, and approximately one-half of all households do not have a separate kitchen. Virtually all houses have no proper ventilation and chimney.<sup>16</sup>

Chhattisgarh is an agricultural state of central India. The population of this state has been ranked as the 16th top most state in India. Remarkably, more than 80% of the population live in rural area and their major work is dependent on agriculture. Nearly everyone uses wood and cow dung for cooking and for other purpose. Due to several adverse effects of indoor air during burning of wood and cow dung, evaluation of the PM and ions concentration is necessary in our study. Therefore, the indoor PM and ions during burning of wood and cow dung in Raipur, Chhattisgarh, central India is investigated. This study tried to examine the PM and ions during burning of indoor materials and that they may be valuable for evaluating future harmful effect.

## Materials and Methods

The indoor burning materials i.e. commonly used different type of wood ( $n = 5$ ) and cow dung ( $n = 1$ ) were selected for the current study. In India, numerous human culture and medicinal practices are influenced by plants i.e. *Tamarindus indica*; it is used as an energy source for cooking, water heating etc. in houses: *Mangifera indica*; it is commonly used to worship the god in Puja or Hawans by burning the stem: *Azadirachta indica*; it is used as an energy source, medicinal plants and to drive away mosquitoes by burning its leaves: *Ipomea nil* and *Acacia arabica* are used as an energy source. The indoor environments (a standard room ( $3 \times 2 \times 3 \text{ m}^3$ ) equipped with one door and one window ( $1 \times 1 \text{ m}^2$ )) i.e. kitchen using homemade clay-stove for wood and cow dung burning was chosen for collection of particulate ( $PM_{10}$ ). A Partisol Model 2300 sequential speciation air sampler (Thermo Fisher Scientific, USA,  $10 \text{ L min}^{-1}$ ) was used for collection of  $PM_{10}$  on 47-mm quartz fiber filters

(Whatmann, QMA) housed in molded filter cassettes, in the indoor environments for the duration of 1 hrs. The air sampler was installed at the ground level and operated in flow rates of  $10 \text{ L min}^{-1}$ . Always at least one blank filter was used to correct the background values. The filters were heated to  $600^\circ\text{C}$  for 6 hrs before exposure so as to reduce the carbon blank values. The filters were weighted by employing the Mettler Toledo balance type - AG245 and placed in the sampler and run for the duration of the burning process. In order to eliminate moisture contents from the loaded filters, it was heated up to  $50^\circ\text{C}$  for 6 hrs. The filters were finally weighted to measure the particulate mass load. The mass of the particulate in the air was calculated by dividing the aerosol mass with the volume of the air passed.<sup>19</sup>

Additional 15 mL of de-ionized water ( $0.054 \mu\text{S cm}^{-1}$ ) was used to extract PM content followed by sonication for 15 min and about 24 hrs heating at  $60^\circ\text{C}$ . After filtration of the extract using tracer filters ( $0.45 \mu\text{m}$  of pore size), about 200  $\mu\text{l}$  of aliquot was added into the ion chromatograph. The ion chromatograph (DX120, Dionex, USA) equipped with anion separation column (AS9-HC,  $250 \times 4 \text{ mm}$ ), cation separation column (CS12A,  $250 \times 4 \text{ mm}$ ) and conductivity detector was used to analyze the ions. The eluents, 9 mM  $\text{Na}_2\text{CO}_3$  ( $1.4 \text{ mL min}^{-1}$ ) and 20 mM methane sulfonic ( $0.8 \text{ mL min}^{-1}$ ) were used for leaching of the anion and cation, respectively. In order to evaluate the soluble ion content of the samples, a standard (AR, E. Merck) was employed for the preparation of the calibration curves. The laboratory blank was used to assess possible contaminations.

The emission flux of  $\text{PM}_{10}$  was determined by burning the materials in a closed chamber ( $0.5 \times 0.5 \times 0.5 \text{ m}^3$ ) consisting of wood equipped with the exhaust fan and UC Davis (USA) portable air sampler (Figure 2). The sampler was mounted on the chamber. Two gram of each material was used for the burning. The burning was carried out till the complete burning of the materials with concurrent collection of the  $\text{PM}_{10}$  over the quartz filter paper ( $47 \text{ mm}$ ). Likewise, the sample blank (i.e. without

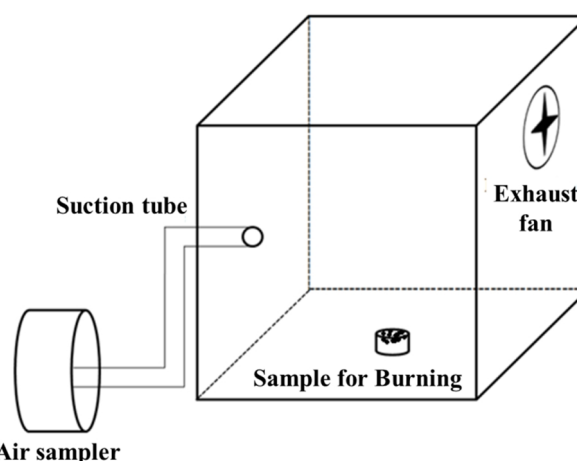
collection on filter) was done for the correction. The  $\text{PM}_{10}$  mass was weighted out, and the flux was evaluated by dividing the  $\text{PM}_{10}$  mass with the amount of the burnt material. The flux for the  $\text{PM}_{10}$  was calculated by using the following equation (1):

$$\text{PM}_{\text{flux}} = \text{PM}_m / W \quad (1)$$

where,  $\text{PM}_m$  and  $W$  represent the mass of  $\text{PM}_{10}$  in the filter and the amount of the burning materials. The flux for the ions associated with the  $\text{PM}_{10}$  was calculated by applying the following equation (2):

$$A_{\text{flux}} = \text{PM}_{\text{flux}} \times F \quad (2)$$

where,  $A_{\text{flux}}$  = Fluxes of ion in the  $\text{PM}_{10}$ ,  $F$  = Ionic fraction in the  $\text{PM}_{10}$ .



**Fig. 2** A closed chamber ( $0.5 \times 0.5 \times 0.5 \text{ m}^3$ ) equipped with the exhaust fan and UC Davis (USA) portable air sampler.

## Results and Discussion

The concentration of  $\text{PM}_{10}$  of the indoor environments during burning of five types of wood ( $n = 5$ ) and cow dung ( $n = 1$ ) is given in Table 1. The PM concentration of wood and cow dung was found to be 17697 and 7768  $\mu\text{g m}^{-3}$ , respectively. In this study, the highest PM concentration was observed in the wood than in cow dung; which indicate that the use of wood as a cooking fuel has been a major factor of increasing PM concentration in indoor environment. Evidently, higher PM concentration of wood and cow dung also demonstrates that they are more harmful fuels compared to the other cleaner fuels like LPG.

Also, the PM concentration of wood and cow dung was found to be several times higher than the guidelines levels i.e.  $150 \mu\text{g m}^{-3}$  for 24 hours. India is a religious country and they burn up wood and cow dung as an energy source as well as also for worship purposes of indoor environments. Among the wood materials

tested, the highest PM concentration was observed with *Mangifera indica* ( $24035 \mu\text{g m}^{-3}$ ) that is used for both purposes i.e. as an energy source as well as for worship; indicating that it is a more harmful material than the other tested materials. *Mangifera indica* is commonly used to worship the god because, Hindu religion in

**Table 1** Concentration of PM<sub>10</sub> and ions in PM<sub>10</sub> in indoor air,  $\mu\text{g m}^{-3}$

Sample No.	Materials	PM <sub>10</sub>	Cl <sup>-</sup>	NO <sub>3</sub> <sup>-</sup>	SO <sub>4</sub> <sup>2-</sup>	NH <sub>4</sub> <sup>+</sup>	Na <sup>+</sup>	K <sup>+</sup>	Mg <sup>2+</sup>	Ca <sup>2+</sup>
W1	<i>Tamarindus indica</i>	19836	6.2	3.2	3.6	0.2	3.8	1.7	0.5	4.0
W2	<i>Mangifera indica</i>	24035	19.8	7.1	8.0	10.4	1.8	16.4	0.5	4.5
W3	<i>Azadirachta indica</i>	15015	8.2	2.7	4.8	0.1	1.8	8.5	0.3	2.6
W4	<i>Ipomea nil</i>	10490	4.4	3.6	5.4	0.2	2.4	2.6	0.6	2.7
W5	<i>Acacia arabica</i>	19107	17.9	6.8	4.7	0.3	5.4	10.7	0.4	3.0
CD6	Cow dung	7768	6.9	2.5	1.9	0.1	1.8	4.3	0.1	0.9

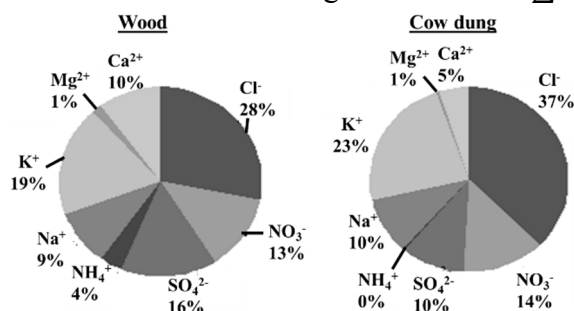
India believes that the god bless good health and wealth to human life by burning stem of this plant.<sup>20</sup> Some studies in India and other developing countries reported high PM concentration during burning of wood and cow dung in the indoor environments.<sup>21-27</sup> The direct comparison of PM between this study and other studies is difficult owing to the use of different wood materials, and different burning and monitoring techniques. Lohani et al. found that PM<sub>10</sub> concentration on wood burning houses is 2 times greater than in LPG and kerosene using houses in Nepal which implies that wood or biomass burning is a more dirty fuel than the LPG and kerosene.<sup>15</sup>

The PM<sub>10</sub> samples derived from burning of wood and cow dung were examined for the determination of eight ions i.e. Cl<sup>-</sup>, NO<sub>3</sub><sup>-</sup>, SO<sub>4</sub><sup>2-</sup>, NH<sub>4</sub><sup>+</sup>, Na<sup>+</sup>, K<sup>+</sup>, Mg<sup>2+</sup> and Ca<sup>2+</sup>, Table 1. The percentage contribution of ions to the sum of eight ions ( $\sum\text{ion}_8$ ) is presented in Figure 3. The mean concentrations of ions for wood ( $38.4 \mu\text{g m}^{-3}$ ) burning were found to be higher than the cow dung ( $18.5 \mu\text{g m}^{-3}$ ). Among the wood materials, the highest  $\sum\text{ion}_8$  concentration was observed with *Mangifera indica* ( $68.5 \mu\text{g m}^{-3}$ ) and the second highest concentration was marked with *Acacia arabica* ( $49.2 \mu\text{g m}^{-3}$ ). Both wood materials are commonly used as an energy source in this study area due to their great availability and *Mangifera indica* is also even used for worship purpose. Likely to the high PM

concentration, *Mangifera indica* was also found to be higher in ionic concentration, which indicates its great harmfulness for indoors than the other wood materials. Almost similar concentration trend of each ion was observed with the wood and cow dung i.e. Cl<sup>-</sup> > K<sup>+</sup> > SO<sub>4</sub><sup>2-</sup> > NO<sub>3</sub><sup>-</sup> > Na<sup>+</sup> > Ca<sup>2+</sup> > NH<sub>4</sub><sup>+</sup> > Mg<sup>2+</sup>. Among all ions, the highest concentration was observed with Cl<sup>-</sup> for both materials i.e. wood ( $11.3 \mu\text{g m}^{-3}$ ) and cow dung ( $6.9 \mu\text{g m}^{-3}$ ). The second dominant ions was found to be K<sup>+</sup> and its concentration for wood and cow dung were  $8.0 \mu\text{g m}^{-3}$  and  $4.3 \mu\text{g m}^{-3}$ , respectively. The reported work also shows that Cl<sup>-</sup> is the most abundant inorganic species in the PM that originated from biomass fuel burning.<sup>28</sup> K<sup>+</sup> is the main component of any kind of biomass burning<sup>8</sup>, and often used as a marker for biomass burning; however the second highest concentration of K<sup>+</sup> ion aside the first was found in this work. The other study conducted in outdoor in Raipur, central India and other Indian sites also reported lower concentration of K<sup>+</sup> ion in the PM.<sup>29,30</sup>

The other work has also reported the higher fraction of K<sup>+</sup> and Cl<sup>-</sup> in PM, which is comparable with our study.<sup>31,32</sup> The concentration of ions was compared with the reported work in India and other region, and it was found that the different trend of ions due to different soil quality, weather, ground water quality etc., can affect the chemical

composition of species of different biomass fuel used in a region. Singh et al. have reported the concentration trend of ions i.e.  $SO_4^{2-} > NO_3^- > Ca^{2+} > Cl^- > K^+ > Na^+ > Mg^{2+} > NH_4^+$ .<sup>33</sup> In  $\sum ion_8$ ,



**Fig.3** Percentage contribution of ions to the  $\sum ion_8$  during burning of materials i.e. wood and cow dung.

the highest anionic contribution was observed with both materials i.e. 56.4% for wood and 61.6% for cow dung. The cationic concentration was found to be 43.6% and 38.8% for wood and cow dung, respectively. The highest anionic contribution was observed with cow dung as a result of high level of  $Cl^-$  and  $NO_3^-$  ions thus indicating that cow dung is a more acidic material than wood.

The ion equilibrium calculations were

measured to observe the acid-base equilibrium of ions particulate in the burning emissions from indoor and are listed in Table 2.<sup>34</sup> Estimation of charge equilibrium between anions and cations was measured by converting the concentrations into ion micro equivalents as the following equation (3):

$$\text{Anion micro equivalents} = Cl^-/35.5 + NO_3^-/62 + SO_4^{2-}/48$$

$$\text{Cation micro equivalents} = NH_4^+/18 + Na^+/23 + K^+/39 + Mg^{2+}/12 + Ca^{2+}/20 \quad (3)$$

The mean particulate equivalent concentration ratio of the  $\sum anion$  to  $\sum cation$  in the wood and cow dung was found to be 0.77 and 0.90, respectively. The equilibrium concentration ratio of both materials shows that the anions were balanced with the cations. Owing to higher anionic concentration of cow dung, it shows that it has more acidic indoor environment than wood burning. Among the wood materials, the *Acacia arabica* was observed to be acidic in nature and their particulate equivalent concentration ratio was 1.01, indicating that its burning can create harmful indoor acidic environment.

**Table 2** Particulate equivalent concentrations of ions,  $\mu Eq$

Sample No.	Cl <sup>-</sup>	NO <sub>3</sub> <sup>-</sup>	SO <sub>4</sub> <sup>2-</sup>	NH <sub>4</sub> <sup>+</sup>	Na <sup>+</sup>	K <sup>+</sup>	Mg <sup>2+</sup>	Ca <sup>2+</sup>	$\sum anion$	$\sum cation$	$\frac{\sum anion}{\sum cation}$
W1	0.17	0.05	0.08	0.01	0.17	0.04	0.04	0.20	0.30	0.46	0.65
W2	0.56	0.11	0.17	0.58	0.08	0.42	0.04	0.23	0.84	1.35	0.62
W3	0.23	0.04	0.10	0.01	0.08	0.22	0.03	0.13	0.37	0.47	0.79
W4	0.12	0.06	0.11	0.01	0.10	0.07	0.05	0.14	0.29	0.37	0.78
W5	0.50	0.11	0.10	0.02	0.23	0.27	0.03	0.15	0.71	0.70	1.01
CD6	0.19	0.04	0.04	0.01	0.08	0.11	0.05	0.05	0.27	0.30	0.90

Majority of the population of our study area live in rural and are commonly using wood and cow dung for household purpose. The construction of numerous kitchen in Chhattisgarh state was poorly ventilated and have shortage of chimney. As a result, the level of toxic contaminants is very high in indoor air resulting in serious health issues.

The emission flux is dependent on two factors i.e. the type of material and their burning condition. The PM<sub>10</sub> emission flux for wood and cow dung was found to be 5.38 and 10.67 g kg<sup>-1</sup>, respectively, Table 3. The emission flux in the case of cow dung is remarkably higher, probably due to its smouldering burning. This smouldering fire enhances the rate of carbon emission and the emission is associated with

particulate matter.<sup>35</sup> Higher PM emission of cow dung may be observed in the pyrolysis process under low temperature<sup>36</sup>, whereas, wood undergoes all three types of burning phases (ignition, flaming and smouldering) than cow dung<sup>37</sup> resulting in a remarkably lower PM emission. Saud et al. reported that the PM emission in wood and cow dung is 4.34 g kg<sup>-1</sup> and 16.26g kg<sup>-1</sup>.<sup>38</sup> Furthermore, Saud et al. measured 4.68 g kg<sup>-1</sup> and 15.68 g kg<sup>-1</sup> PM emission for wood and cow dung, respectively. The above study was found to be very close to our results.<sup>39</sup> The emission flux, 5.38 g kg<sup>-1</sup> obtained from wood burning is consistency with reported results as 5.66 g kg<sup>-1</sup>.<sup>40,41</sup> However, another study found the same fact that cow dung (5.37 g kg<sup>-1</sup>) has higher emission flux than

wood (1.69 g kg<sup>-1</sup>), owing to its smouldering fire. The emission flux of four different woods which ranged from 1.12-2.89 g kg<sup>-1</sup> was reported in Portugal and is lower compared to the present investigation, due to different chemical composition of wood.<sup>42</sup>

The emission flux of  $\Sigma_{ion8}$  for cow dung (52.84 g kg<sup>-1</sup>) was found to be 8-folds higher

**Table 3** Emission flux of PM<sub>10</sub> and ions, g kg<sup>-1</sup>

Sample No.	PM <sub>10</sub>	Cl <sup>-</sup>	NO <sub>3</sub> <sup>-</sup>	SO <sub>4</sub> <sup>2-</sup>	NH <sub>4</sub> <sup>+</sup>	Na <sup>+</sup>	K <sup>+</sup>	Mg <sup>2+</sup>	Ca <sup>2+</sup>
W1	3.12	0.81	0.31	0.31	0.03	0.62	0.22	0.06	0.62
W2	1.62	1.01	0.32	0.32	0.49	0.10	0.81	0.02	0.16
W3	1.95	0.16	0.06	0.10	0.002	0.04	0.16	0.006	0.58
W4	10.78	2.26	2.16	3.23	0.11	1.08	1.08	0.32	0.12
W5	9.41	4.52	1.88	0.94	0.09	0.94	2.82	0.09	0.75
CD6	10.67	21.35	7.47	5.34	0.21	5.34	10.67	0.32	2.14

than wood (5.94 g kg<sup>-1</sup>), which may be due to their burning in smouldering fire, Table 3. Higher ionic flux is associated with the PM emission. The ions emission flux obtained from cow dung is higher and similar as compared to other reports.<sup>34,39</sup> The anion emission trend of the current study has also been compared with other reports<sup>43-45</sup>, and it was found that the soil quality, weather, ground water quality etc. can affect the chemical composition of species of different wood used in a region<sup>39</sup> and, therefore, the difference was found with the emissions.

The correlation concentration between PM<sub>10</sub> and ions of wood burning is shown in Table 4.

The reported correlation values >0.70, are highlighted in bold (see Table 4). The analysis of correlation values of K<sup>+</sup> and Cl<sup>-</sup> (r = 0.84) illustrates their origin from same source. The reported study also indicated higher correlation between K<sup>+</sup> and Cl<sup>-</sup> by biosmoke and derived K<sup>+</sup> should exist in the form of KCl.<sup>46</sup> However, another reported work demonstrates that K<sup>+</sup> is associated with the particulate that are generated from biomass burning.<sup>47</sup> Therefore, it can be concluded that they were generated from the same sources (wood burning) and should exist in the form of KCl.

**Table 4** Correlation coefficient of PM<sub>10</sub> and ions from burning of wood materials

	PM <sub>10</sub>	Cl <sup>-</sup>	NO <sub>3</sub> <sup>-</sup>	SO <sub>4</sub> <sup>2-</sup>	NH <sub>4</sub> <sup>+</sup>	Na <sup>+</sup>	K <sup>+</sup>	Mg <sup>2+</sup>	Ca <sup>2+</sup>
PM <sub>10</sub>	1								
Cl <sup>-</sup>	0.58	1							
NO <sub>3</sub> <sup>-</sup>	0.43	0.88	1						
SO <sub>4</sub> <sup>2-</sup>	0.15	0.38	0.21	1					
NH <sub>4</sub> <sup>+</sup>	0.48	0.46	0.57	0.84	1				
Na <sup>+</sup>	0.02	0.05	0.01	0.29	0.19	1			
K <sup>+</sup>	0.42	0.84	0.14	0.6	0.61	0.03	1		
Mg <sup>2+</sup>	0.02	0.05	0.1	0.05	0.04	0.01	0.11	1	
Ca <sup>2+</sup>	0.73	0.21	0.99	0.21	0.57	0.01	0.14	0.1	1

## Conclusion

In the present study, approximately 80% of the population of Chhattisgarh state live in rural area and depend upon biomass i.e. wood and cow dung for cooking and other purpose. The wood like *Mangifera indica* is used for both energy source as well as to worship the god in indoors and it emits higher concentration of particulate which results in adverse health effect such as respiratory diseases, eyes infections, etc. Majority of the people use *Acacia arabica* in the form of energy sources

due to its greater availability and its acidic appearance in nature. However, the higher emission flux was observed with cow dung than wood. Hence, to improve the quality of indoor air, some recommended options should apply in future such as cleaner source of energy (LPG), sufficient ventilation to exhaust toxic gases and installation of improved models. The outcome of our investigation is for people to be aware of the adverse effects of indoor emission and also to use cleaner fuel and to prefer improved model for better indoor environment.

## Acknowledgement

Authors are thankful to the Head of the Department, School of Studies in Chemistry, Pt. Ravishanker Shukla University, Raipur, Chhattisgarh for providing lab facility. We also give sincere acknowledgement to Atmospheric Pollution Laboratory, Applied Physics Department, Miguel Hernandez University, Avda de la Universidad S/N, 03202 Elche, Spain, for providing financial support to visit the Research Centre.

## Reference

1. Kulshreshtha P, Khare M. Indoor exploratory analysis of gaseous pollutants and respirable particulate matter at residential homes of Delhi, India. *Atmos Pollut Res* 2011; 2(3): 337-350.
2. Mishra VK, Retherford RD, Smith KR. Biomass cooking fuels and prevalence of tuberculosis in India. *Int J Infect Dis* 1999; 3(3):119-129.
3. Smith KR. Fuel combustion, air pollution exposure, and health: the situation in developing countries. *Annu Rev Energy Environ* 1993; 18: 529-566.
4. Bruce N, Perez-Padilla R, Albalak R. Indoor air pollution in developing countries: a major environmental and public health challenge. *Bull World Health Organ* 2000; 78(9): 1078-1092.
5. NSSO (National Sample Survey Organization) Household Consumption Expenditure in India, 2007-08. Ministry of Statistics & Programme Implementation. Government of India. 2010.
6. Banerjee R. et al. Rural energy, renewable energy, energy efficiency and demand side management. IIT Bombay. Draft paper submitted to integrated energy policy committee. Planning commission. October 2004.
7. Andreae MO, Merlet MO. Emission of trace gases and aerosols from biomass burning. *Global Biogeochem Cycles* 2001; 15(4):955-966.
8. Kulshreshtha UC, Sekar R, Vairamani M, Jain M, Sarkar AK, Parashar DC, et al. Signatures of biomass burning over Indian ocean during INDOEX IGAC Symposium, Jan. 21-23, Bangkok, Thailand. 2001.
9. Lobert JM, Scharffe DH, Hao WM, Kuhlbusch TA, Seuwen R, Warneck P, et al. Experimental evaluation of biomass burning emissions: nitrogen and carbon containing compounds. (In J.S. Levine (Ed.), *Global biomass burning: atmospheric, climatic and biospheric implications*. Cambridge, Mass.). 1991. 289-304
10. Bingemer H. et al. Biomass burning as a source of sulfur to the atmosphere. *Eos Transactions, American Geophysical Union*. 72. 86, 1991.
11. Andreae MO, Atlas E, Cachier H, Cofer III WR, Harris GW, Helas G, et al. Trace gas and aerosol emissions from savanna fires. (In J.S. Levine (Ed.), *Biomass burning and global change, Remote Sensing, Modeling and Inventory Development, and Biomass Burning in Africa*, vol.1, (pp. 278-295). MIT Press, Cambridge, Mass.). 1996.
12. Wang Y, Zhuang GS, Zhang XY, Huang K, Xu C, Tang AH, et al. The Ion chemistry, Seasonal cycle and sources of PM<sub>2.5</sub> and TSP aerosol in Shanghai. *Atmos Environ* 2006; 40(16): 2935-2952.
13. Fullerton DG, Bruce N, Gordon SB. Indoor air pollution from biomass fuel smoke is a major health concern in the developing world. *Trans R Soc Trop Med Hyg* 2008; 102(9): 843-851.
14. Gall ET, Carter EM, Earnest CM, Stephens B. Indoor air pollution in developing countries: research and implementation needs for improvements in global public health. *Am J Public Health*. 2013; 103(4): 67-72.
15. Lohani SP. Biomass as a source of household energy and indoor air pollution in Nepal. *Iran J Energy Environ* 2011; 2(1):74-78.
16. Prasad R, Singh A, Garg R, Hosmane GB. Biomass fuel exposure and respiratory diseases in India. *Biosci Trends* 2012; 6(5):219-228.
17. Zhang JJ, Smith KR. Household air pollution from coal and biomass fuels in China: measurements, health impacts, and interventions. *Environ Health Persp* 2007; 115(6):848-855.
18. Patel N, Okocha B, Narayan S, Sheth M. Indoor air pollution from burning biomass & child health. *Int J Sci Res* 2013; 2(1): 492-506.
19. Behera SN, Sharma M. Investigating the potential role of ammonia in ion chemistry of fine particulate matter formation for an urban environment. *Sci Total Environ* 2010; 408(17):3569-3575.
20. Sharma V, Joshi BD. Role of sacred plants in religion and health-care system of local people of Almora district of Uttarakhand state (India). *Academ Arena* 2010; 2(6):19-22.
21. Raiyani CV, Shah SH, Desai NM, Venkaiah K, Patel JS, Parikh DJ, et al. Characterization and problems of indoor pollution due to cooking stove smoke. *Atmos Environ* 1993; 27(11):1643-1655.

22. Balakrishnan K, Sambandam S, Ramaswamy P, Mehta S, Smith KR. Exposure assessment for respirable particulates associated with household fuel use in rural districts of Andhra Pradesh, India. *J Expo Anal Environ Epidemiol* 2004; 14:S14-S25.
23. Colbeck I, Nasir ZA, Ali Z. Indoor levels of particulate pollution in urban & rural environments in Pakistan. In *Proceeding of the European Aerosol Conference, Karlsruhe, Germany. 2009 September 6-11.*
24. Ansari FA, Khan AH, Patel DK, Siddiqui H, Sharma S, Ashquin M, et al. Indoor exposure to respirable particulate matter and particulate-phase PAHs in rural homes in north India. *Environ Monit Assess* 2010; 170 (1-4):491-497.
25. Mondal NK, Konar S, Banerjee A, Datta JK. A comparative assessment of status of indoor air pollution of few selected families of rural and urban area of Burdwan town and its adjoining area. *Int J Environ Sci* 2011; 1(5):736-743.
26. Massey DD, Kulshrestha A, Taneja A. Particulate matter concentrations and their related metal toxicity in rural residential environment of semi-arid region of India. *Atmos Environ* 2013; 67:278-286.
27. Semmens EO, Noonan CW, Allen RW, Weiler EC, Ward TJ. Indoor particulate matter in rural, wood stove heated homes. *Environ Res* 2015; 138: 93–100.
28. Reid JS, Koppmann R, Eck TF, Eleuterio DP. A review of biomass burning emissions part II: intensive physical properties of biomass burning particles. *Atmos Chem Phys* 2005; 5(3):799-825.
29. Deshmukh DK, Deb MK, Mkomla SL. Size Distribution and Seasonal Variation of Size-segregated Particulate Matter in the Ambient Air of Raipur City, India. *Air Qual Atmos Health* 2013; 6(1):259-276.
30. Rengarajan R, Sudheer AK, Sarin MM. Wintertime PM<sub>2.5</sub> and PM Carbonaceous and Inorganic Constituents from Urban Site in Western India. *Atmos Res* 2011; 102(4):420–431.
31. Huboyo HS, Tohno S, Lestari P, Mizohata A, Okumura M. Characteristics of indoor air pollution in rural mountainous and rural coastal communities in Indonesia. *Atmos Environ* 2014; 82:343-350.
32. Park SS, Soo YS, Bae MS, Schauer JJ. Size distribution of water-soluble components in particulate matter emitted from biomass burning. *Atmos Environ* 2013; 73:62-72.
33. Singh S, Gupta GP, Bablu Kumar, Kulshrestha UC. Comparative study of indoor air pollution using traditional and improved cooking stoves in rural households of Northern India. *Energy Sustain Develop* 2014; 19:1–6.
34. Sen A, Mandal TK, Sharma SK, Saxena M, Gupta NC, Gautam R, et al. Chemical properties of emission from biomass fuels used in the rural sector of the western region of India. *Atmos Environ* 2014; 99:411-424.
35. Koppmann R, von Czapiewski K, Reid JS. A review of biomass burning emissions, part I: gaseous emissions of carbon monoxide, methane, volatile organic compounds, and nitrogen containing compounds. *Atmos Chem Phys Discuss* 2005; 5:10455-10516.
36. Parashar DC, Gadi R, Mandal TK, Mitra AP. Carbonaceous aerosol emissions from India. *Atmos Environ* 2005; 39(40):7861-7871.
37. Venkataraman C, Negi G, Sardar SB, Rastogi R. Size distribution of polycyclic aromatic hydrocarbons in aerosol emissions from biofuel combustion. *J Aero Sci* 2002; 33(3):503–518.
38. Saud T, Mandal TK, Gadi R, Singh DP, Sharma SK, Saxena M, et al. Emission estimates of particulate matter (PM) and trace gases (SO<sub>2</sub>, NO and NO<sub>2</sub>) from biomass fuels used in rural sector of Indo-Gangetic Plain, India. *Atmos Environ* 2011; 45(32):5913–5923.
39. Saud T, Saxena M, Singh DP, Saraswati, Dahiya M, Sharma SK, et al. Spatial variation of chemical constituents from the burning of commonly used biomass fuels in rural areas of the Indo-Gangetic Plain (IGP), India. *Atmos Environ* 2013; 71:158-169.
40. Cao GL, Wang FC, Wang YQ. Emission inventory of TSP, PM<sub>10</sub> and PM<sub>2.5</sub> emission from biomass burning in China. *J Process Eng* 2004; 700-704.
41. EPD. Guide for compiling atmospheric pollutant emission inventory for biomass burning. Environmental Protection Department. 2014.
42. Gonçalves C, Alves C, Evtyugina M, Mirante F, Pio C, Caseiro A, et al. Characterisation of PM<sub>10</sub> emissions from woodstove combustion of common woods grown in Portugal. *Atmos Environ* 2010; 44(35):4474-4480.
43. Salam A, Hasan M, Begum BA, Begum M, Biswas SK. Chemical characterization of biomass burning deposits from cooking stoves in Bangladesh. *Biomass Bioenergy* 2013; 52:122-130.



44. Deka P, Hoque RR. Incremental effect of festive biomass burning on wintertime PM10 in Brahmaputra valley of Northeast India. *Atmos Res* 2014; 143:380-391.
45. Deka P, Hoque RR. Chemical characterization of biomass fuel smoke particles of rural kitchens of South Asia. *Atmos Environ* 2015; 108:125-132.
46. Lee CT, Chuang MT, Lin NH, Wang JL, Sheu GR, Chang SC, et al. The enhancement of PM2.5 mass and water-soluble ions of biosmoke transported from Southeast Asia over the Mountain Lulin site in Taiwan. *Atmos Environ* 2011; 45(32):5784–5794.
47. Khare P, Goel A, Patel D, Behari J. Chemical characterization of rainwater at a developing urban habitat of Northern India. *Atmos Res* 2004; 69(3-4):135–145.